

RESEARCH ARTICLE

Potentially toxic element contamination and risk assessment of borehole water within a landfill in the Nnewi metropolis

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Abstract: Borehole water has been used as a vital source of water for many communities. The pollution of these boreholes by potentially toxic elements using unlined solid waste dumpsites has posed a significant risk to the populace living around the dumpsite. This study investigates the environmental impact of potentially toxic elements on borehole water within the vicinity of an unlined dumpsite in the Nnewi metropolis. The upstream and downstream samples (16 each) were collected in wet and dry seasons. The potentially toxic elements analysis wasperformed using the methods of the American Public Health Association (APHA) under the required conditions. The quality of the boreholes was assessed using the World Health Organisation (WHO) acceptable limits for drinking water. The risk assessment was estimated for carcinogenic and non-carcinogenic risks using ingestion and dermal routes. The results show that the borehole water wascontaminated with potentially toxic elements through leachate infiltration, which exceeded the WHO permissible limits for drinking water at both locations and seasons. The upstream borehole samples were more contaminated than the downstream samples for both seasons, due to their proximity to the pollution source. The hazard indices of the ingestion and dermal routes showed that the borehole water poses serious cancer and non-cancer health risks for both locations. The results revealed that children are more susceptible to cancer and non-carcinogenic health threats than adults for both locations and seasons. The pollution indices of borehole water for wet season (9.028 and 5.728) and dry season (7.107 and 5.328) for upstream and downstream samples respectively, were polluted and the pollution was higher in the wet season. The borehole water samples were unsuitable for drinking water and should be treated before use.

Keywords: potentially toxic elements, borehole, contamination, Nnewi, health risk

1 Introduction

Groundwater has always been a major source of water for the residents of Nnewi in Anambra State. The populace normally depends on groundwater for drinking and other domestic purposes. Groundwater quality can be altered because of human activities, which can change the physicochemical characteristics of the water [1,2]. Human activities such as unsanitary landfills, poor soakaway systems and indiscriminate dumping of refuse can have negative consequences on the quality of the underground water [3]. Unsanitary landfills are the most common waste disposal practiseused in Anambra State, Nigeria [4–6].

The increase in population growth, urbanisation and industrialisation has also increased the volume of waste generated in Anambra State [7]. Poor allocations for waste management have negatively impacted waste management activities [7]. This method of waste practice has led to environmental pollution, which can pose a severe health risk to the populace [8–15]. The unregulated leachates coming from this unlined dumpsite tend to infiltrate the soil and thereby pollute the soil and groundwater due to the migration of pollutants [16–19].

Leachates from these dumpsites according to researchers constitute a major source of potentially toxic elements pollution to the groundwater and its environment [20–24]. Wastes such as electronic products, paint waste and automobile batteries deposited in the refuse dump, tend to increase the volume of potentially toxic elements in dumpsites, which can be toxic to the environment [25]. Open burning of wastes using unlined dumpsite is a common practise usually done to reduce the quantity of waste [6]. The by-products left after the burning can pose a risk to the underground water and public health [25]. Potentially toxic elements are adsorbed into soils and water bodies, which when exposed to the human body, results ina severe threat to human health [12, 26]. Recently, the contamination and potential risk of underground water have been investigated [27–29].

Some heavy metals such as Pb and Hg, are toxic and can induce cancer in the human body when they exceed their threshold limits [22, 26, 29, 30]. Exposure to potentially toxic elements toxicity can cause brain damage, dermatitis, anaemia and death in humans [26, 31].

Health-associated risks caused by the use of unlined dumpsites call for a complete assessment of the effect of unlined waste dumpsites on the environment [6]. Many researchers have reported groundwater pollution through leachates from dumpsites [25, 32–34]. However, few studies have been conducted potentially toxic elements pollution of underground water from landfill leachates in the Nnewi metropolis [6,35]. The previous studies on the underground water quality in Nnewi only focused on the levels of potentially toxic elements and pollution indices of the water and did not evaluate the risk assessment involved in the use of the water by the residents.

Therefore, in this research work, the impact of solid waste leachates on potentially toxic elements contamination of groundwater quality was studied around the Okpunoeze dumpsite in the Nnewi metropolis. Cancer and non–cancer health risks were also estimated using dermal and ingestion routes on the populace. The metal pollution index of the borehole water was also determined. The findings of this work will be a valuable tool for policy makers on better ways to prevent underground water pollution using improved waste disposal methods.

2 Materials and methods

2.1 Sampling area

The sampling area (Figure 1) is located around the OkpunoezeOtolo dumpsite in Nnewi, Anambra State. Nnewi is located within the tropical rainforest region of Nigeria [36]. Nnewi is a major town in Anambra State known for its manufacturing activities, which has led to the rapid growth of the city [36, 37]. Open dumpsites are the predominant waste disposal method practised in the area [1]. The dumping site is located at the following latitude of $6^{\circ}00'43.4''N$ and longitude of $6^{\circ}54'28.2''E$ [37]. Two distinct seasons are observed, dry and wet. Nnewi has an annual rainfall of about 2000 mm [38]. Wastes disposed of in the landfill are predominantly solid waste from industries, hospitals, markets, workshops and households, which are located around the dumpsite [1].



Figure 1 Map of sampling areas

The Nnewi metropolis lies beneath Nanka Formation, consisting of loose and friable sands [35]. The lithology of the area is composed of sandstones, which are porous and allow for infiltrating leachates into the aquiferous system [35]. The average depth to the water table in the study area is about 110 m and the average static water level is 120 m [39].

2.2 Sampling of borehole water samples

Borehole water samples (16 each) were collected during the morning hours (9 - 10 am) from the upstream and downstream locations for 4 months (May-August 2018) for the wet season,

and 3 months (December 2018 - February 2019) in the dry season [1]. The downstream samples were collected 685 - 935 m away from the upstream samples, which are located around the dumpsite [37]. Properly washed and rinsed glass sample bottles were used for sample collection from the selected locations. The sampled borehole water was homogenised to form a composite sample [37]. Acidification of the samples was done with 10% HNO₃ [1]. The water samples were brought in an ice chest, before being stored at 4° C in the refrigerator [37].

2.3 Chemicals and reagents

All chemicals used were of high analytical reagent grade, which were purchased from Sigma-Aldrich, USA.

2.4 Sample preparation and potentially toxic elements analysis

The potentially toxic elements analyses of the samples were carried out using standard methods [40]. 100 mL of the borehole water samples were transferred into a 250 mL glass beaker. Concentrated HNO₃(6 mL) was added and it was heated using a hot plate until the volume was reduced to 20mL. The mixture was allowed to cool and filtered. It was made up to the 100 mL mark using deionized water. The samples were aspirated into the oxidizing airacetylene flame using an Agilent 240-FS Atomic Absorption Spectrophotometer. The sensitivity for 1% absorption was observed after the samples were aspirated. The potentially toxic elements analyses was analysed for nickel (Ni), iron (Fe), cadmium (Cd), chromium (Cr), lead (Pb), arsenic (As),cobalt (Co), manganese (Mn), zinc (Zn) and copper (Cu).

A standard solution containing 1000 ppm of 2% HNO₃ was used to prepare the spiking experiments and calibrationstandards [33]. Two working standards for each potentially toxic elements were prepared from these standards. A standard calibration curve was obtained by running a prepared standard solution of each potentially toxic element [33].

2.5 Quality control

The blank samples and certified reference materials (CRM) were analyzed to ensure the accuracy, reliability and reproducibility of the results [33]. The results were found within $\pm 5\%$ of the certified values and a recovery rate between 93.4 and 101.2%. The analyses were performed in triplicate with the mean values recorded [37].

2.6 Health risk assessment

The potential health risk of the borehole water on the residents was assessed using ingestion and dermal exposure assessment, which was expressed using Equations 1 and 2 [41].

$$CDD_{ing} = \frac{C \times IR \times ED \times EF}{BW \times AT} \times CF$$
(1)

$$CDD_{derm} = \frac{C \times SA \times SAF \times DAF \times ED \times EF}{BW \times AT} \times CF$$
(2)

Where CDD_{ing} and CDD_{derm} represent the chronic daily dose of ingestion and dermal contact (mg/kg/day). The exposure parameters used are presented in Table 1. For non-carcinogenic risks, the hazard quotients (HQs) for ingestion and dermal contact were valuated using Equations 3 and 4 [42,43]. The hazard index (HI) was evaluated using Equation 5. The HQ or HI < 1, signifies no risk, while the HQ or HI l≥ signifies risk.

 Table 1
 Parameters for health risk assessment

Parameters	Unit	Adult	Child	Reference
Concentration of metals (C)	mg/L	-	-	-
Ingestion rate (IR)	L/day	2.5	0.78	[41]
Exposure duration (ED)	years	20	6	[45]
Exposure frequency (EF)	days/year	365	365	[45]
Average time (AT) non-cancer	days	7300	2190	[41]
Average time (AT) cancer	days	25550	25550	
Body weight (BW)	kg	70	15	[45]
Surface area (SA)	cm^2	6032	2373	[45]
Skin adherence factor (SAF)	mg/cm ²	0.07	0.2	[46]
Dermal absorption factor (DAF)	-	0.001	0.001	[41]
Conversion factor (CF)	-	1E-06	1E-06	[41]

$$HQ = \frac{CDD_{ing}}{RfD}$$
(3)

$$HQ = \frac{CDD_{derm}}{RfD}$$
(4)

$$HI = \sum HQs \tag{5}$$

The cancer risks (CR) and HI in the boreholewater were estimated using Equations 6, 7 and 8 [44].

$$CR = CDD_{ing} \times SF$$
 (6)

$$CR = CDD_{derm} \times SF$$
 (7)

$$HI = \sum CDDs$$
(8)

Table 2 shows the reference doses and slope factors for the potentially toxic elements.

Table 2 Reference doses and cancer slope factors for non-carcinogenic and carcinogenic risks

	Reference of	loses (RfD)	Slope fa	Reference	
Potentially toxic elements (mg/L)	Dermal	Ingestion	Dermal	Ingestion	Reference
Zinc	0.06	0.3	-	-	[47]
Arsenic	0.000123	0.0003	3.66	1.5	[47]
Lead	0.000524	0.0014	-	0.0085	[47]
Iron	0.7	0.7	-	-	[47]
Nickel	0.00054	0.02	4.25	0.91	[47]
Manganese	0.014	0.014	-	-	[47]
Chromium	0.003	0.003	2	0.5	[47]
Cobalt	0.02	0.03	-	-	[47]
Cadmium	0.000025	0.001	-	6.3	[47]
Copper	0.012	0.04	-	-	[47]

2.7 Metal Pollution Index (MPI)

MPI represents the sum of the ratio between the analysed parameters and their corresponding national standard values [48]. The rating is a value between 0 and 1 [48].

$$MPI = \sum_{i=1}^{n} \left[\frac{Ci}{MAC} \right]$$
(9)

Where Ci = mean concentration, MAC = maximum allowable concentration.

2.8 Statistical evaluation

Correlation analyses between upstream and downstream samples on the levels of potentially toxic elements in the borehole water were performed. The relationships between the parameters were evaluated using a hierarchical cluster dendogram. SPSS version 23 software was used for the analysis.

3 Results and discussion

3.1 Potentially toxic elements analyses

Table 3 shows the potentially toxic elements parameters of the borehole water samples. The borehole samples showed different levels of potentially toxic elements characteristics in both study areas. The values of the mean potentially toxic elements parameters were compared with WHO standard limits (Figure 2).

Table 3	Potentially	toxic el	ements l	level ir	n the	borehole	water
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		Wet S	Season				[<mark>49</mark>]		
Metals	Metals Upstream		Downstream		Ups	stream	Down		
	Range	Mean	Range	Mean	Range	Mean	Range	Mean	
Zn	1.38-1.58	$1.478 {\pm} 0.08$	1.14-1.37	$1.245 {\pm}~0.09$	1.18-1.31	$1.23 {\pm} 0.07$	1.12-1.24	$1.187{\pm}0.06$	3
As	0.01 - 0.02	$0.013 {\pm} 0.01$	ND-0.01	$0.008 {\pm}~0.00$	0.01 - 0.01	$0.01 {\pm} 0.00$	ND-0.01	$0.007 {\pm} 0.00$	0.01
Pb	ND-0.02	$0.018 {\pm} 0.00$	ND-0.02	0.01 ± 0.01	0.01 - 0.02	$0.013 {\pm} 0.01$	ND-0.01	$0.007 {\pm} 0.00$	0.01
Fe	0.28-0.36	$0.308 {\pm} 0.04$	0.22-0.25	$0.235 {\pm} 0.01$	0.24-0.31	$0.267 {\pm} 0.04$	0.17-0.24	$0.207 {\pm} 0.04$	0.3
Ni	0.04 - 0.07	$0.05 {\pm} 0.02$	ND-0.04	$0.023 {\pm} 0.02$	0.05 - 0.06	$0.053 {\pm} 0.01$	0.03-0.05	$0.037 {\pm} 0.01$	0.07
Mn	0.37-0.45	$0.4 {\pm} 0.04$	0.10-0.34	$0.265 {\pm}~0.11$	0.32-0.38	$0.347 {\pm} 0.03$	0.22-0.33	$0.277 {\pm} 0.06$	0.4
Cr	0.03-0.05	$0.043 {\pm} 0.01$	0.01 - 0.04	0.03 ± 0.01	0.03-0.04	$0.033 {\pm} 0.01$	0.01-0.03	$0.023 {\pm} 0.01$	0.05
Co	0.03-0.05	$0.04{\pm}0.01$	0.02 - 0.04	0.03 ± 0.01	0.03-0.04	$0.037 {\pm} 0.01$	0.02-0.03	$0.023 {\pm} 0.01$	-
Cd	ND-0.01	$0.005 {\pm} 0.00$	ND-0.01	$0.003 {\pm} 0.00$	ND-0.01	$0.003 {\pm} 0.00$	ND-0.01	$0.003 {\pm} 0.00$	0.003
Cu	0.37-0.48	$0.433{\pm}0.05$	0.32-0.48	$0.383 {\pm} 0.07$	0.41-0.47	$0.443 {\pm} 0.03$	0.28-0.37	$0.32{\pm}0.05$	2



Figure 2 Seasonal variation of mean potentially toxic elements parameters for both study areas

The mean concentration of Zn ranged from 1.23 mg/L - 1.478 mg/L for upstream samples for both seasons. In the downstream samples, the mean values ranged from 1.187 - 1.245 mg/L for both seasons. The Zn values for the upstream sample were higher than the downstream values, possibly due to infiltration of the leachates from the dumpsite, which were closer to the upstream locations. The Zn levels were below the WHO permissible limit (3 mg/L) for drinking water. The values were lower than 1.79 mg/L and 10.8 mg/L which werereported [25, 34]. The Zn values obtained from this study were higher than the values obtained [33].

The As values for both seasons ranged from 0.01 - 0.013 mg/L for the upstream samples, while the downstream samples ranged from 0.007 - 0.008 mg/L for both seasons. The As values in the upstream samples were equal to or greater than the WHO maximum allowable limit of 0.01 mg/L for both seasons. Higher values of arsenic higher than the study samples werereported [29, 34, 50]. The values of As reported in this study were similar [25, 32].

The mean values of Pb in the upstream location ranged from 0.013 to 0.018 mg/L for both seasons, while the downstream values ranged from 0.01 - 0.007 mg/L. The values were equal to or greater than the WHO threshold limit of 0.01 mg/L, except for the downstream value in the dry season, which was lower than the WHO limit. The high values of Pb could be attributed to the disposal of lead batteries, pipes, and paints at the refuse dump [33, 34]. The values obtained from this study were similar [29, 33]. High values of Pb higher than this study was obtained [25, 32, 33].

The Fe values for the upstream values ranged from 0.267 - 0.308 mg/L for both seasons, while the downstream values ranged from 0.207 - 0.237 mg/L respectively, which were above the permissible limit of 0.30 mg/L. The mean value for the upstream samples during the wet season was the only mean value that exceeded the WHO permissible limits. High values of Fe in water can be attributed to the oxidation of metal wastes in the dumpsites [35]. The upstream samples were higher than the downstream samples for both seasons. The values in this study were lower than obtained reports [32, 33, 51]. The values in this study were higher than the values obtained by [29, 34].

The mean values of Ni for the upstream samples for both seasons ranged from 0.05 - 0.053 mg/L. The downstream samples ranged from 0.023 - 0.037 mg/L for both locations. The values were lower than the threshold limit of 0.07 mg/L for both locations in both seasons. The upstream location values were higher than the downstream locations for both seasons. The Ni values in this study were similar [32]. The Ni values from this study were higher than the reported work [29].

The mean values of Mn for the upstream samples ranged from 0.347 - 0.4 mg/L, while they ranged from 0265 - 0.277 mg/L for the downstream samples for both seasons. The upstream sample value in the wet season was the only Mn value equal to the WHO threshold, while the rest were within the limit. The values from this study were higher than the obtained value [32] and were lower than the values [52].

The mean values of Cr for the upstream samples ranged from 0.033 - 0.043 mg/L, while the downstream values ranged from 0.023 - 0.03 mg/L for both seasons. These values were below the permissible limit of the WHO. The values in this study were similar [33]. The mean values of chromium obtained in this study were lower than the results obtained [25, 29, 51].

The mean values of Co for the upstream samples for both seasons ranged from 0.03 - 0.04 mg/L, while the downstream sample values ranged from 0.023 - 0.03 mg/L. These values were lower than the results obtained [53, 54].

The mean values for Cd ranged from 0.003 - 0.005 mg/L for the upstream samples and 0.003 0.003 mg/L for the downstream samples for both seasons. The Cd values were either equal to or greater than the threshold limit by the WHO for both locations and seasons. Cadmium could be linked with the dispersal of potentially toxic elements produced from electronic wastes disposed of in the dumpsite [34]. Similar values of Cd were reported [33, 34]. The values were higher than the reported work [29] and lower than obtained work [25, 52].

The mean values for Cu in the upstream sample ranged from 0.443 - 0.443 mg/L for both seasons, while they ranged from 0.32 - 0.383 mg/L in the downstream samples in both seasons. These values were below the WHO permissible limits of 2 mg/L. The values were higher than the values obtained [25, 32, 33, 35] and lower than the value obtained [34].

The order of maximum concentrations followed this order for the wet season: Zn > Cu>Mn>Fe>Ni > Cr > Co>Pb>As > Cd for the upstream samples, and Zn > Cu>Mn>Fe>Cr>Co>Ni > Pb>As>Cd for the downstream samples. The dry season followed this order: Zn > Co>Mn>Fe > Ni > Co > Cr>Pb>As > Cd for the upstream samples and Zn > Co>Mn>Fe > Ni > Co>Cr>Pb>As > Cd for the downstream samples.

The correlation values of the potentially toxic elements parameters (Table 4) for the borehole water samples indicated a strong positive correlation between upstream and downstream samples during the wet season (r = 0.998, P = 0.000) and dry season (r = 0.995, P = 0.000). The correlation study values between the wet and dry season upstream values (r = 0.998, P = 0.000) and downstream values (r = -0.999, P = 0.000) showed astrong positivelinear relationship [37]. The strong correlation observed in the study areas indicates that the pollution are of a similar source [37]. The pollutant source was linked to anthropogenic activities such as the discharge of leachates from the unlined dumpsite [37]. The p-values are all less than 0.05 (p < 0.05), which implies that the potentially toxic elements characteristics are significant, and are dependent on the extent of accumulation of pollutants in the borehole water samples in both seasons [37].

Therefore, the study suggests that the levels of potentially toxic elements obtained were attributed to leachate percolation through the unlined refuse dump. The upstream sample's results were predominantly higher than the results of the downstream samples for both locations and seasons. This is due to the infiltration of pollutants from the leachates because of its proximity to the dumpsite [44].

Table 4	Pearsons correlation	tion betw	veen st	tudy are	ea param	eters acros	ss both sea	asons	

	Upstream samples (Wet season)	Downstream samples (Wet season)	Upstream samples (Dry season)	Downstream samples (Dry season)
Upstream samples (Wet season)	1			
Downstream samples (Wet season)	0.998**	1		
Upstream samples (Dry season)	0.998**	0.997**	1	
Downstream samples (Dry season)	0.999**	0.999**	0.995**	1

Note: ** Correlation is significant at the 0.01 level (2-tailed)

3.2 Non-carcinogenic risk via ingestion route

The non-carcinogenic risk values through the ingestion route are shown in Table 5. The hazard index values for the upstream locations were greater than those for the downstream locations for both seasons. This trend shows that the child's HI values were predominantly greater than the adult's HI, which implies that the children are at a higher risk than the adult [32,34]. The HI values for the upstream location were greater than the downstream location because of its proximity to the dumpsite.

3.3 Carcinogenic risk through ingestion routes

The carcinogenic risk values lower than 1.0E-06 are a negligible risk, while values above 1.0E-04 represent a potential threat likely to humans [41]. Table 6 shows the carcinogenic risk values through the ingestion route. The cancer risk values of lead for adults and children were below the carcinogenic threshold, while the cancer risk values for Cr, Cd, As and Ni were higher than the threshold limit indicating cancer threat. Generally, the HI values for both age categories were above the threshold value of 1.0E-04, which indicates a potential cancer risk to

		Ad	ults		Children				
Parameters	Wet season Dr			Dry season Wet s		season	Dry season		
	Upstream	Downstream	Upstream	Downstream	Upstream	Downstream	Upstream	Downstream	
Zn	0.1476	0.1244	0.1229	0.1186	0.2457	0.2069	0.2044	0.1973	
As	1.2985	0.7991	0.9989	0.6992	2.1607	1.3297	1.6621	1.1635	
Pb	0.3853	0.214	0.2783	0.5244	0.6411	0.3562	0.463	0.2493	
Fe	0.0132	0.0101	0.0114	0.0089	0.0219	0.0167	0.019	0.0147	
Ni	0.0749	0.0345	0.0794	0.0554	0.1247	0.0573	0.1321	0.0922	
Mn	0.8562	0.5672	0.7427	0.5929	1.4247	0.9438	1.2359	0.9866	
Cr	0.4295	0.2997	0.3296	0.0005	0.7147	0.4986	0.5485	0.3823	
Co	0.04	0.03	0.037	0.023	0.0665	0.0499	0.0615	0.0382	
Cd	0.1498	0.0899	0.0899	0.0899	0.2493	0.1496	0.1496	0.1496	
Cu	0.3244	0.2869	0.3319	0.2397	0.5398	0.4774	0.5522	0.3989	
HI	3.7194	2.4558	3.022	2.3525	6.1891	4.0861	5.0283	3.6726	

 Table 5
 Hazard quotient of non-carcinogenic risks through the ingestion route

the population. The child's cancer risk values were higher than those of the adult, indicating that the children are more exposed to cancer threats [34, 55]. Moreover, for both age categories, Ni has a higher potential cancer threat compared to other potentially toxic elements. The HI for the upstream samples was greater than that for the downstream locations due to their closeness to the dumpsite.

oute

		Ad	ults		Children				
Parameters	Wet season		Dry season		Wet season		Dry season		
	Upstream	Downstream	Upstream	Downstream	Upstream	Downstream	Upstream	Downstream	
As	2E-04	1E-04	1E-04	9E-05	8E-05	5E-05	6E-05	4E-05	
Pb	1E-06	7E-07	9E-07	5E-07	7E-07	4E-07	5E-07	3E-07	
Ni	4E-04	2E-04	4E-04	3E-04	2E-04	9E-05	2E-04	1E-04	
Cr	2E-04	1E-04	1E-04	1E-04	9E-05	6E-05	7E-05	5E-05	
Cd	3E-04	2E-04	2E-04	2E-04	1E-04	8E-05	8E-05	8E-05	
HI	1E-03	6E-04	8E-04	6E-04	5E-04	3E-04	4E-04	3E-04	

3.4 Non-carcinogenic health risk via dermal contact

Table 7 shows the non-carcinogenic health risk values through dermal contact. The other potentially toxic elements' HQ values were all lower than 1. The HI values for the adult were higher than 1 for the upstream location during the wet season. The child HI values were all higher than 1, with the upstream location greater than the downstream samples. Children are more exposed to non-carcinogenic related risks than adults because of higher HI values [34].

 Table 7
 Hazard quotient of non-carcinogenic risks through the dermal route

		Adults				Children				
Parameters	Wet	Wet season		Dry season		Wet season		season		
	Upstream	Downstream	Upstream	Downstream	Upstream	Downstream	Upstream	Downstream		
Zn	0.1247	0.105	0.1038	0.1001	0.7474	0.6295	0.622	0.6002		
As	0.5349	0.3292	0.4115	0.288	3.2066	1.9733	2.4666	1.7267		
Pb	0.1739	0.0966	0.1256	0.0676	1.0422	0.579	0.7527	0.4053		
Fe	0.0022	0.0017	0.0019	0.0015	0.0133	0.0102	0.0116	0.009		
Ni	0.4686	0.2156	0.4967	0.3468	2.8092	1.2922	2.9778	2.0788		
Mn	0.1446	0.0958	0.1254	0.1001	0.8668	0.5743	0.752	0.6003		
Cr	0.0725	0.0506	0.0557	0.0388	0.4349	0.3034	0.3337	0.2326		
Co	0.0101	0.0076	0.0094	0.0058	0.0607	0.0455	0.0561	0.0349		
Cd	1.0122	0.6073	0.6073	0.6073	6.0679	3.6408	3.6408	3.6408		
Cu	0.1826	0.1615	0.1868	0.135	1.0948	0.9683	1.12	0.8091		
HI	2.7263	1.6709	2.1241	1.691	16.3438	10.0165	12.7333	10.1377		

3.5 Carcinogenic risk through dermal contact

The carcinogenic risk values via dermal contact with potentially toxic elements are shown in Table 8. The carcinogenic risk values through dermal routes were above the threshold value of 1.0E-04 for the potentially toxic elements calculated. The HI values were all above the threshold limit. The upstream values were greater than the downstream values because of

leachate infiltration into the dumpsite [32]. The HI for the children was higher than that of adults, which makes them more susceptible to cancer risk through dermal contact. A similar trend was reported [34].

		Ad	ults		Children			
Parameters	Wet season		Dry season		Wet season		Dry season	
	Upstream	Downstream	Upstream	Downstream	Upstream	Downstream	Upstream	Downstream
As	7E-05	4E-05	5E-05	4E-05	1E-04	8E-05	1E-04	7E-05
Ni	3E-04	1E-04	3E-04	2E-04	6E-04	3E-04	6E-04	4E-04
Cr	1E-04	9E-05	1E-04	7E-05	2E-04	2E-04	2E-04	1E-04
HI	5E-04	3E-04	5E-04	3E-04	9E-04	5E-04	9E-04	6E-04

 Table 8
 Carcinogenic risks through the dermal route

3.6 Metal pollution index

The potentially toxic elements pollution index of the borehole water samples for the wet and dry seasons is shown in Table 9. The values showed that the pollution index was beyond the critical value of 1, and hence, cannot be used for specific purposes without any special treatment. The potentially toxic elements pollution index for the wet season was 9.028 and 5.782, while the dry seasonwas 7.107 and 5.328 for the upstream and downstream respectively. The potentially toxic elements pollution index for the wet season was higher than the dry season in both locations because of the percolation of leachates from the refuse dump aided by rainfall. The upstream samples pollution index was higher than the downstream samples, which was attributed to the infiltration of more leachates due to proximity to the refuse dump. Therefore, the borehole water samples should not be used for domestic purposes without special treatment.

 Table 9
 Metal pollution index of borehole samples

	Wet	season	Dry season		
Parameters	Upstream	Downstream	Upstream	Downstream	
Zinc	0.493	0.415	0.41	0.396	
Arsenic	1.25	0.8	1	0.7	
Lead	1.8	1	1.3	0.7	
Iron	1.027	0.783	0.89	0.69	
Nickel	0.714	0.329	0.757	0.529	
Manganese	1	0.663	0.868	0.693	
Chromium	0.86	0.6	0.66	0.46	
Cobalt	-	-	-	-	
Cadmium	1.667	1	1	1	
Copper	0.217	0.192	0.222	0.16	
$\sum MPI$	9.028	5.782	7.107	5.328	

3.7 Cluster analysis of potentially toxic elements

The potentially toxic elements parameters that had similar characteristics were grouped using dendogram hierarchical cluster [7]. The hierarchical cluster dendrogram (HCA) showed four distinct clusters (Figure 3). Ni was grouped into the first cluster, while Zn and Cd were grouped into the fourth cluster. Co and Cu were grouped into the second cluster. The As, Pb, Fe, Cr and Mn were grouped into the third cluster. The cluster showed that Ni is not influenced by any other elements assessed, while As is influenced by Pb, Fe, Cr and Mn. The observed clustering could be linked to the anthropogenic sources of the potentially toxic elements in the leachates of the refuse dump [34].

4 Conclusion

The results of the study area showed that the upstream and downstream boreholes were contaminated with different concentrations of potentially toxic elements through the infiltration of the leachates from the unlined dumpsite. The results revealed that Cd was above the WHO limit across both locations. The Mn and Fe values for the upstream samples in the wet season were above the WHO permissible limit. The As mean values in the upstream samples in both seasons were above the WHO limit. The Pb values were above the WHO threshold limit, except for the downstream sample values in the dry season. The overall results obtained showed that both borehole water samples do not meet the WHO standard for drinking water. The upstream sample's potentially toxic elements' levels were greater than that of the downstream samples, because of leachate infiltration linked to their proximity to the dumpsite. The wet



Figure 3 Hierarchical cluster dendrogram of the potentially toxic elements (heavy metals)

season concentrations were higher than those during the dry season because of the runoff of leachates from the unlined dumpsite. Carcinogenic and non-carcinogenic risks through ingestion and dermal contact with potentially toxic elements pose serious health-related risks to the population, of which children are more vulnerable than adults. The pollution indices showed that the borehole water samples were polluted and should be treated before use. Routine monitoring of the borehole water around the dumpsite area should be encouraged to curtail health-related risks from exposure to potentially toxic elements toxicity. The government should adopt an efficient waste management system that involves constructing sanitary dumpsite that will prevent leachate infiltration from the environment.

Author contribution

All authors have read and approved the manuscript. CCA: Writing- Original draft preparation, Conceptualization, Methodology; PAC: Reviewing and Editing, Data curation; HOA: Investigation, Supervision and Validation; VCE: Methodology, Data curation, Software; HOC: Visualization, Conceptualization. The final manuscript was read and approved by all authors.

Availability of data

The datasets generated during and/or analysed during the current study are available from the corresponding author on reasonable request.

Compliance with ethical standards

All authors have read, understood, and have complied as applicable with the statement on "Ethical responsibilities of Authors" as found in the Instructions for Authors.

Disclosure statement

No potential conflict of interest is reported by the authors.

References

- Aralu CC, Okoye PAC and Akpomie KG. Toxicological risk evaluation of polycyclic aromatic hydrocarbons in boreholes located within unsanitary dumpsite in Nnewi, Anambra State, Nigeria. International Journal of Environmental Analytical Chemistry, 2021. https://doi.org/10.1080/03067319.2021.1916485
- [2] Alam R, Ahmed Z, Seefat SM, *et al.* Assessment of surface water quality around a landfill using multivariate statistical method, Sylhet, Bangladesh. Environmental Nanotechnology Monitoring & Management, 2020, **15**(4): 100422. https://doi.org/10.1016/j.enmm.2020.100422

- [3] Ubechu BO, Ikoro DO, Irefin MO, *et al.* Assessment of physicochemical characteristics of groundwater around an unlined landfill in Aba, Southeastern Nigeria. Research Journal of Environmental & Earth Sciences, 2021, 7(10): 76-82.
- [4] Aralu CC, Okoye PAC, Akpomie KG, et al. Levels of polycyclic aromatic hydrocarbons in leachates from unlined dumpsite of Agu-AwkaAnambra State. International Journal of Environmental Analytical Chemistry, 2021. https://doi.org/10.1080/03067319.2021.1993842
- [5] Aralu CC, Okoye PAC, Ezugwu JA, et al. Assessment of Water Collected from Boreholes Within Unsanitary Landfill in Anambra State. European Journal of Applied Science, 20021, 9(6):405-12. https://doi.org/10.14738/aivp.96.11365
- [6] Momoh SO, Mgbemena CO, Agbadua SA, et al. The physicochemical effect of leachates on groundwater within OkpunoezeUmudim dumpsite Nnewi, Anambra State Nigeria. Daffodil International University Journal of Science and Technology, 2013, 8(2):25-32.
- [7] Aralu CC, Okoye PAC, Akpomie KG, et al. Polycyclic aromatic hydrocarbons in soil situated around solid waste dumpsite in Awka, Nigeria. Toxin reviews, 2022(1/2): 41. https://doi.org/10.1080/15569543.2021.2022700
- [8] Mansourri G and Madani M. Examination of the level of heavy metals in wastewater of Bandar Abbas Wastewater Treatment Plant. Open Journal of Ecology, 2016, 6: 55-61. https://doi.org/10.4236/oje.2016.62006
- [9] Okalebo SE, Opata GP and Mwasi BN. An analysis of the household solid waste generation patterns and prevailing management practices in Eldoret Town, Kenya. International Journal of Agricultural Policy and Research, 2014, 2(2): 76.
- [10] Wokhe TB. Heavy Metals Pollution of Water and Sediment in Mada River, Nigeria. Journal of Scientific Research and Reports, 2015, 6(2): 157-164. https://doi.org/10.9734/JSRR/2015/14803
- [11] Kusin FM, Azani N, Hasan S, et al. Distribution of heavy metals and metalloid in surface sediments of heavily-mined area for bauxite ore in Pengerang, Malaysia and associated risk assessment. Catena, 2018, 165: 454-464.

https://doi.org/10.1016/j.catena.2018.02.029

- [12] Zhao K, Fu W, Qiu Q, et al. Spatial patterns of potentially hazardous metals in paddy soils in a typical electrical waste dismantling area and their pollution characteristics. Geoderma, 2019, 337: 453-462. https://doi.org/10.1016/j.geoderma.2018.10.004
- [13] Kosheleva NE, Vlasov DV, Korlyakov ID, et al. Contamination of urban soils with heavy metals in Moscow as affected by building development. Science of the Total Environment, 2018, 636: 854-863. https://doi.org/10.1016/j.scitotenv.2018.04.308
- [14] Aralu CC and Okoye PAC. Impact of solid waste leachates on soil and edible plants within unlined dumpsite in Awka, Anambra State. American Journal of Chemistry, 2020, 10(1): 11-18. https://doi.org/10.1080/03067319.2021.1993842
- [15] Aralu CC and Okoye PAC. Assessment of heavy metals levels in soil and vegetables in the vicinity of unlined waste dumpsite in Nnewi, Anambra State Nigeria. Journal of Chemical Society of Nigeria, 2020, 45(4): 687-696.
 https://doi.org/10.046602/joan.y45i4.402
 - https://doi.org/10.46602/jcsn.v45i4.493
- [16] Aralu CC, Okoye PAC, Abugu HO, et al. Toxicity and distribution of polycyclic aromatic hydrocarbons in leachates from an unlined dumpsite in Nnewi, Nigeria. International Journal of Environmental Analytical Chemistry, 2022. https://doi.org/10.1080/03067319.2022.2140415
- [17] Bagul VR, Shinde DN, Chavan RP, *et al.* New perspective on heavy metal pollution of water. Journal of Chemical & Pharmaceutical Research, 2015, **7**(12): 700-705.
- [18] Kumar R, Tripathi RM and Gupta AK. Seasonal Variation of heavy metal concentration in water of River Yamuna, Allahabad, Uttar Pradesh, India. International Journal of Current Microbiology & Applied Sciences, 2014, 3(7): 945-949.
- [19] Majolagbe AO, Adeyi AA and Osibanjo O. Vulnerability assessment of groundwater pollution in the vicinity of an active dumpsite (Olusosun), Lagos, Nigeria. ChemInt, 2016, 2: 232-241.
- [20] Omaka ON, Aghamelu OP, Ike-Amadi CA, et al. Assessment of the quality of groundwater from different parts of southeastern Nigeria for potable use. Environmental Earth Sciences, 2017, 76(9):344. https://doi.org/10.1007/s12665-017-6680-z
- [21] Chen M, Qin X, Zeng G and Li J. Impacts of human activity modes and climate on heavy metal "spread" in groundwater are biased. Chemosphere, 2016, 152: 439-445. https://doi.org/10.1016/j.chemosphere.2016.03.046
- [22] Adimalla N. Groundwater quality for drinking and irrigation purposes and potential health risks assessment: A case study from semi-arid region of south India. Exposure and Health, 2018. https://doi.org/10.1007/s12403-018-0288-8
- [23] Adimalla N. Spatial distribution, exposure, and potential health risk assessment from nitrate in drinking water from semi-arid region of South India. Human and Ecological Risk Assessment: An International Journal, 2019. https://doi.org/10.1080/10807039.2018.1508329
- [24] Mazhari SA, Bajestani ARM, Hatefi F, *et al.* Soil geochemistry as a tool for the origin investigation and environmental evaluation of urban parks in Mashhad city, NE of Iran. Environmental Earth Sciences, 2018, 77(13): 492. https://doi.org/10.1007/s12665-018-7684-z

[25] Maiti SK, De S, Hazra T, *et al.* Characterization of leachate and its impact on surface and groundwater quality of a closed dumpsite-a case study of Dhapa, Kolkata, India. Procedia Environmental Sciences, 2016, **35**: 391-399.

https://doi.org/10.1016/j.proenv.2016.07.019

- [26] Khademi H, Gabarrón M, Abbaspour A, et al. Environmental impact assessment of industrial activities on heavy metals distribution in street dust and soil. Chemosphere, 2019, 217: 695-705. https://doi.org/10.1016/j.chemosphere.2018.11.045
- [27] Ogbuene EB, Odinkonigbo UL, Nebo AN, *et al.* Open solid waste leachate output and monthly rainfall variation in Ugwuaji, Enugu Nigeria. International Journal of Development and Sustainability, 2018, 7(3): 986-992.
- [28] Michaela EI, Odohb AO, Chukwurac EI, *et al.* Heavy Metal and Microbial Load Properties of Dumpsite Leachate: Case Study of Onitsha Dumpsite, South-East, Nigeria. Journal of Analytical Toxicology, 2018, 1: 1-6.
- [29] Ali SKA and Ahmad A. Analysis of chemical and heavy metal concentrations of leachates and impact on groundwater quality near Dhapa dumping ground, Kolkata. Asian Profile, 2019, 47(1): 19-37.
- [30] Ukah BU, Egbueri JC, Unigwe CO, et al. Extent of heavy metals pollution and health risk assessment of groundwater in a densely populated industrial area, Lagos, Nigeria. International Journal of Water Resources, 2019, 3: 291-303. https://doi.org/10.1007/s42108-019-00039-3
- [31] Sankhla MS, Kumari M, Nandan M, et al. Heavy metals contamination in water and their hazardous effect on human health-A Review. International Journal of Current Microbiology and Applied Sciences, 2016, 5(10):759-766. https://doi.org/10.20546/ijcmas.2016.510.082
- [32] Chakraborty S and Kumar RN. Assessment of groundwater quality at a MSW landfill site using standard and AHP based water quality index: a case study from Ranchi, Jharkhand, India. Environmental Monitoring and Assessment, 2016, 188(6): 335. https://doi.org/10.1007/s10661-016-5336-x
- [33] Boateng TK, Opoku F and Akoto O. Heavy metal contamination assessment of groundwater quality: a case study of Otilandfll site, Kumasi. Applied Water Science, 2019, 9: 33. https://doi.org/10.1007/s13201-019-0915-y
- [34] Laniyan TA and Adewumi AJ. Heavy Metal Pollution in Groundwater in Southwestern Nigeria. Journal of Health & Pollution, 2019, 9: 22. https://doi.org/10.5696/2156-9614-9.24.191210
- [35] Egbueri JC. Assessment of the quality of groundwaters proximal to dumpsites in Awka and Nnewi metropolises: a comparative approach. International Journal of Energy & Water Resources, 2018. https://doi.org/10.1007/s42108-018-0004-1
- [36] Obeta MC. Industrial water supply in Nnewi urban area of Anambra State, South Eastern Nigeria. Journal of Geography, Environment and Earth Science International, 2015, 2(1): 12-23. https://doi.org/10.9734/JGEESI/2015/14538
- [37] Aralu CC, Okoye PAC, Abugu HO, *et al.* Pollution and water quality index of boreholes within unlined waste dumpsite in Nnewi, Nigeria. Discovery Water, 2022, 2: 14. https://doi.org/10.1007/s43832-022-00023-9
- [38] Ekenta OE, Okoro BU and Ezeabasili ACC. Hydrogeological Characteristics and Groundwater Quality Analysis for Selected Boreholes in Ogbaru Local Government Area, Anambra State, Nigeria. American Scientific Research Journal for Engineering, Technology, and Sciences, 2015, 14(2): 198-210.
- [39] Nfor BN, Olobaniyi SB and Ogala JE. Extent and distribution of groundwater resources in parts of Anambra state, Southeastern Nigeria. Journal of Applied Sciences and EnvironmentalManagement, 2007, 11(2): 215-221. https://doi.org/10.4314/jasem.v11i2.55050
- [40] American Public Health Association (APHA). Standard Methods for the Examination of Water and Wastewater, 21st ed.; APHA, AWWA, WEF: Washington, DC, USA, 2005.
- [41] USEPA. Supplemental guidance for developing soil screening levels for superfund sites. Washington: U.S. Environmental Protection Agency, Office of Emergency and Remedial Response, 2002.
- [42] Deng Y, Jiang L, Xu L, et al. Spatial distribution and risk assessment of heavy metals in contaminated paddy fields-A case study in Xiangtan City, southern China. Ecotoxicology and Environmental Safety, 2019, **171**: 281-289.

https://doi.org/10.1016/j.ecoenv.2018.12.060

- [43] Chen H, Teng Y, Lu S, et al. Contamination features and health risk of soil heavy metals in China. Science of the Total Environment, 2015, 512: 143-153. https://doi.org/10.1016/j.scitotenv.2015.01.025
- [44] Eze VC, Owukeme VI and Ebere EC. Pollution status, ecological and human health risks of heavy metals in soil from some selected active dumpsites in Southeastern, Nigeria using energy dispersive X-ray spectrometer. International Journal of Environmental Analytical Chemistry, 2020, 1-22.. https://doi.org/10.1080/03067319.2020.1772778
- [45] USEPA. Exposure Factors Handbook 2011 Edition (Final).National Center for Environmental Assessment, Office of Research and Development. Washington D.C., 2011. https://cfoub.epa.gov
- [46] US EPA (United States Environmental Protection Agency). Risk assessment guidance for superfund volume I: human health evaluation manual (Part E, Supplementary guidance for dermal risk assessment) final. EPA/540/R/99/005. Washington, DC: US EPA, 2004, 1-156.

- [47] USEPA. Risk assessment guidance for superfund. Human health evaluation manual (Part A) (Vol. 1). Washington, DC: Office of Emergency and Remedial Response, 1989.
- [48] Tamasi G and Cini R. Heavy metals in drinking waters from Mount Amiata. Possible risks from arsenic for public health in the province of Siena. Science of the Total Environment, 2004, 327: 41-51. https://doi.org/10.1016/j.scitotenv.2003.10.011
- [49] WHO. Guidelines for drinking-water quality: fourth edition incorporating the first addendum, 2017, 1-631.
- [50] Longe EO and Enekwechi LO. Investigation on potential groundwater impacts and influence of local hydrogeology on natural attenuation of leachate at a municipal landfill.International Journal of Environmental Science and Technology, 2007, 4(1): 133-140. https://doi.org/10.1007/BF03325971
- [51] Alam P, Sharholy M and Ahmad K. A Study on the Landfill Leachate and Its Impact on Groundwater Quality of Ghazipur Area, New Delhi, India. Recent Developments in Waste Management, Lecture Notes in Civil Engineering, 2020, 57: 345-358. https://doi.org/10.1007/978-981-15-0990-2_27
- [52] Fernández DS, Puchulu ME and Georgieff SM. Identification and assessment of water pollution as a consequence of a leachate plume migration from a municipal landfill site (Tucumán, Argentina). Environ Geochem Health, 2014, 36(3): 489-503. https://doi.org/10.1007/s10653-013-9576-1
- [53] Okeke DO, Ifemeje JC and Eze VC. Determination of the levels of heavy metals and physicochemical properties of borehole water within selected mining sites in Ebonyi state, Nigeria. International Journal of ChemBiol Science, 2021, 3(2): 5-10. https://doi.org/10.25082/HE.2022.01.003
- [54] Michael EI. Evaluation of leachate contamination index obtained from dumpsite in Onitsha, Nigeria. Journal of Environmental and analytical toxicology, 2017, 7(6): 525. https://doi.org/10.4172/2161-0525.1000525
- [55] Eze V, Ndife C and Muogbo M. Carcinogenic and Non-carcinogenic Health Risk Assessment of Heavy Metals in Njaba River, Imo State, Nigeria. Brazilian Journal of Analytical Chemistry, 2021, 8(33): 57-70.

https://doi.org/10.30744/brjac.2179-3425